



Research Article

ISSN : 0975-7384
CODEN(USA) : JCPRC5

Heavy metal ions removal from water using modified zeolite

Peng Zhang^{1*}, Wenjie Ding¹, Yanjun Zhang², Kanghai Dai¹ and Wei Liu¹

¹School of Civil Engineering, Hunan University of Science and Technology, Xiangtan, Hunan, China

²School of Life Science, Hunan University of Science and Technology, Xiangtan, Hunan, China

ABSTRACT

Zeolite was modified by HNO_3 solution and NaOH solution in order to improve the adsorption performance. The characterization of the modified zeolite by scanning electron microscope (SEM) and infrared spectroscopy (IR) was displayed. In order to obtain the optimum conditions for adsorption of Sb^{3+} and Mn^{2+} , resulting in a better removal efficiency, parameters such as the concentration of HNO_3 and NaOH , dosage of modified zeolite, adsorption time, initial concentration and adsorption temperature were examined. Results showed that the zeolite modified by 0.8 mol/L HNO_3 had better adsorption rate for Sb^{3+} , the zeolite modified by 1.4 mol/L NaOH had better adsorption rate for Mn^{2+} , and the highest adsorption rate of Sb^{3+} and Mn^{2+} were 89.1% and 99.6%, respectively. Isotherm process was well predicted by Freundlich model when adsorption Sb^{3+} by using HNO_3 modified zeolite, and the Langmuir model fitting the equation well when adsorption Mn^{2+} by using NaOH modified zeolite. Kinetic process was predicted well by pseudo second-order model when adsorption Sb^{3+} by using HNO_3 modified zeolite and adsorption Mn^{2+} by using NaOH modified zeolite.

Key words: Adsorption, Antimony ion, Manganese ion, Modified Zeolite, Isotherms, Kinetics

INTRODUCTION

Due to the large-scale continuous mining and mineral separation for long-term, it has resulted in numbers of heavy metals entering into mine soil [1-2]. Upon rainfall, the heavy metal pollutants from contaminated soil is easy to dissociate and transform in soil-water interface, forming a pollution stream of heavy metal along with the rainfall infiltration and transporting by surface runoff, leading to a heavy metal pollution to the water environment [3-4]. Now heavy metal pollution in water environment and the harmful to human are very outstanding [5-6]. Once the heavy metals into the environment, they could not degradation, and there will be persist presence in the water environment. Numerous methods such as physical, chemical and biological processes including adsorption, biosorption, precipitation, ion-exchange, reverse osmosis, filtration and other membrane separations are employed to treat the heavy metal contaminated water [7]. However, adsorption has proven to be economical and efficient for removing heavy metals [8]. Several adsorbents such as zeolite, silica, and active carbon could be used in the water treatment. Zeolite as an kind of adsorbent material has lots of advantages, such as wide range of sources of raw materials, cheap and good adsorption effect, and it is widely used as adsorbents, ion exchangers and catalysts which could also be used for drying gases, purification and sewage treatment [9-10]. The present work aimed to used HNO_3 and NaOH as modifier in order to improve the adsorption ability of zeolite, then the modified zeolite was prepared and characterized by the SEM and IR techniques. In order to obtain better adsorption performance, some parameters such as the independent variables, modified zeolite, contact time, initial concentration and temperature, were investigated, combined with the study on adsorption isotherm and adsorption kinetics.

EXPERIMENTAL SECTION

2.1 Material and instruments

All the reagents used in this study were analytical grade. The hydrochloric acid (HCl) and sodiumhydroxide (NaOH) were obtained from Hunan Zhuzhou Chemical Industry. The antimonous chloride was obtained from Sinopharm Chemical Reagent Co., Ltd. Manganese chloride was obtained from Tianjin Fengchuan Chemical Reagent Technologies Co., Ltd. Natural zeolite was obtained from Crystal Water Treatment Material Co., Ltd. Henan Branch. The equipments used in the experimental setup were as follows:

- Scanning electron microscope (SEM) (JSM-6380LV, JEOL Company, Japan)
- Fourier Transformed Infra Red (FT-IR) Spectrophotometer (Niclet 6700, Niclet instrument Company, USA)
- Atomic Absorption Spectrometer (AA-7001, East & West Analysis Instrument Co., Ltd., China)
- Water bath oscillator (THZ-82, Jintan Ronghua Instrument Manufacturing Co., Ltd., China)

2.2 Pretreated zeolite and modification

First, the zeolite was placed in a beaker which was added 100mL ultra pure water for immersing 2h, filtered, and then washed three times and dried in a vacuum at 105°C for several days. Finally immersing the zeolite with 0.8 mol/L HNO₃ solution and 1.4 mol/L NaOH solution. In 50 °C water bath heating 4h, pouring the supernatant, washing with deionized water for 3 times, dried in a vacuum at 85°C for use.

2.3 Batch adsorption studies

Erlenmeyer flasks (250mL) containing 1.2 g of modified zeolite in 100 mL of antimony trichloride solutions were employed. Modified zeolite were dosed under the rotary shaker at 150 rpm for 4 h, the mixtures were then centrifuged at 3000 r/min for 5 min. The residual Sb³⁺ and Mn²⁺ concentration in the solution was then determined using atomic absorption spectrometer. The adsorption capacity was calculated using the following equation:

$$Q_e = \frac{C_0 - C_e}{W} V$$

where Q_e is the adsorption capacity (mg/g), C₀ is the initial DMP concentration (mg/L), C_e is the final DMP concentration (mg/L), W is the weight of adsorbent used (g) and V is the volume of DMP solution (L). The experiments were carried out in triplicates to obtain average results.

2.4 Isotherm study

Isotherm study was conducted using batch equilibrium experiments. Absorbent 1.2 g was added into every conical flask with 100 mL of aqueous solution containing the concentration solutions of Sb³⁺ between 5.0 and 25.0mg/L, and the concentration solutions of Mn²⁺ between 50.0 and 250.0mg/L. The pH of solution need not to adjust. The mixtures were covered and oscillation in a shaker at a constant speed of 150 r/min and temperature of room temperature. The mixtures were then centrifuged at 3000 r/min for 5 min. The residual Sb³⁺ and Mn²⁺ concentration in the solution was then determined using atomic absorption spectrometer.

2.5 Kinetic study

Batch kinetic experiments were performed by mixing a fixed amount of absorbent (1.2 g) with 100 mL Sb³⁺ solution of Sb³⁺ concentration at 10mg/L for 30-240 min and the Mn²⁺ concentration at 100.0mg/L for 30-270 min. The pH of solution need not to adjust. The mixtures were covered and oscillation in a shaker at a constant speed of 150 r/min and temperature of room temperature. After a certain period of oscillation, the entire suspension was taken from a conical flask and solid-liquid separation was achieved by centrifugation at 3000 r/min for 5 min. The residual Sb³⁺ and Mn²⁺ concentration in the solution was then determined using atomic absorption spectrometer.

RESULTS AND DISCUSSION

3.1 Characterization of the product

The SEM images of modified zeolite was depicted in Figure 1. It can be seen that the surface morphology of the zeolite has been modified after HNO₃ and NaOH treatment. There was a significant changes from the original sizes, shapes crumb structure, transformed into the pore structure. So compared with natural zeolites, it have a stronger adsorption capacity. The Infrared Spectroscopy was the natural zeolite, the product of zeolite modified by HNO₃ and NaOH.

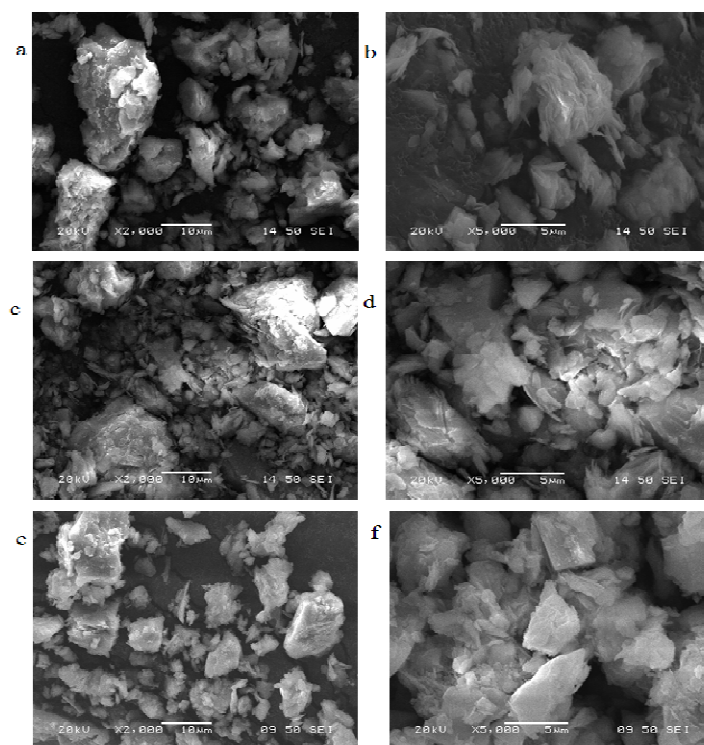
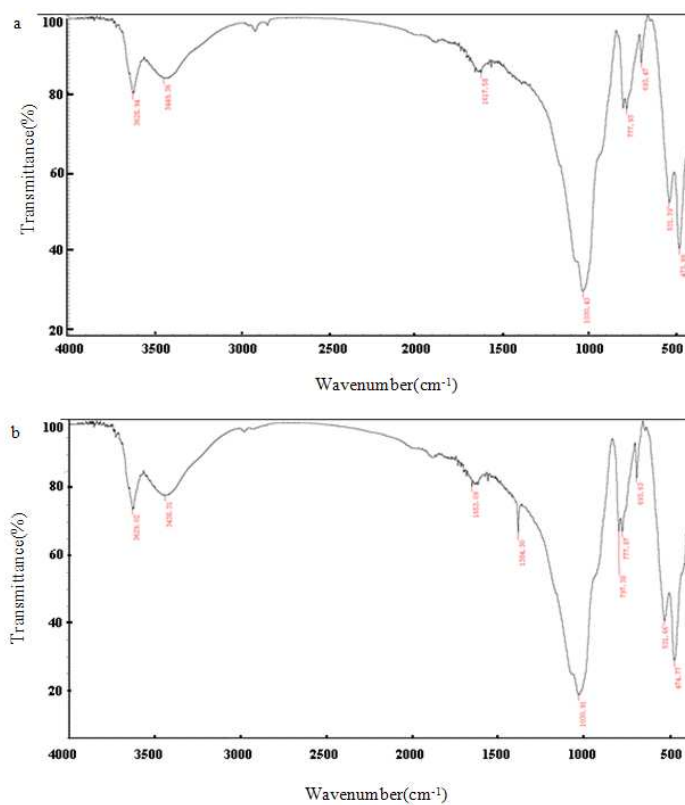


Figure 1. SEM images of Natural zeolite(a, b), Modified zeolite by HNO₃(c,d), Modified by NaOH(e,f)



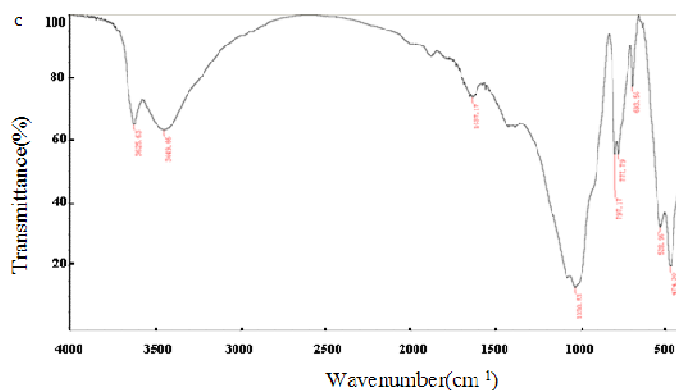


Figure 2. IR spectroscopy of Natural zeolite(a), modified zeolite by HNO_3 (b), modified zeolite by NaOH (c)

3.2 Single influential factor analysis

3.2.1 The selection of modifiers

The effect of the kind of modifier was investigated. HNO_3 , NaOH , and EDTA were used, respectively. The results as shown (Figure 3) that zeolite modified by HNO_3 was had the best adsorption of Sb^{3+} and the zeolite modified by NaOH was had the best adsorption of Mn^{2+} . In order to fixed the concentration of the HNO_3 as modifier. The concentration of HNO_3 ranged from 0.2 up to 1.0 mol/L was selected to investigate the adsorption effect. As the increased of concentration of HNO_3 ranged from 0.2 g to 0.8 mol/L, the removal rate of Sb^{3+} increased from 63.1% to 82.7%, and when the concentration increased to 1.0 mg/L, the adsorption decreased to 62.8%. So 0.8 mol/L was chosen as an optimum concentration. And in order to fixed the concentration of the NaOH as modifier. The concentration of NaOH ranged from 0.2 up to 2.0 mol/L was selected to investigate the adsorption effect. As the increased of concentration of NaOH ranged from 0.2 g to 1.4 mol/L, the removal rate of Mn^{2+} increased from 75.1% to 99.7%, and when the concentration increased to 2.0 mol/L, the adsorption decreased. So 1.4 mol/L was chosen as an optimum concentration.

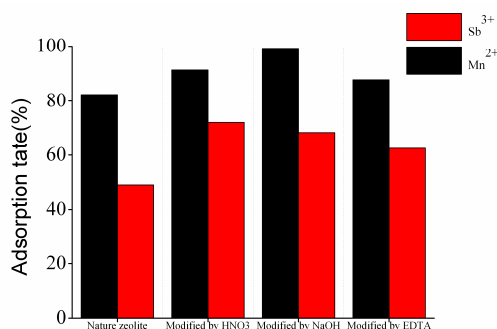


Figure 3. Effect of modifier on adsorption

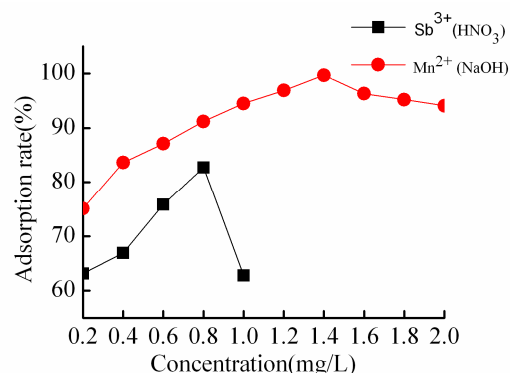


Figure 4. Modifier concentration on adsorption

3.2.2 Effect of modified zeolite dosage on adsorption capacity and adsorption rate

The effect of the dosage of modified zeolite adsorption was investigated. The dosage of modified zeolite ranged from 0.6 up to 1.4 g was selected to investigate the adsorption effect (Figure 5). As shown in when the dosage was increased, the adsorption capacity and removal rate was increased, that because with the increase of the amount of modified zeolite, the total surface area of the modified zeolite increased, which resulting in an increase of the adsorbent effective adsorption sites. As the increased of HNO_3 modified zeolite dosage ranged from 0.6 g to 1.2 g, the removal rate of Sb^{3+} increased from 69.2% to 88.5%, and when the dosage increased to 1.4 g, the adsorption decreased to 87.6%. So 1.2 g was chosen as an optimum dosage. As the increased of NaOH modified zeolite dosage ranged from 0.6 g to 1.0 g, the removal rate of Mn^{2+} increased from 97.6% to 99.3%, and when the dosage increased to 1.4 g, the adsorption increased smoothly. So 1.0 g was chosen as an optimum dosage.

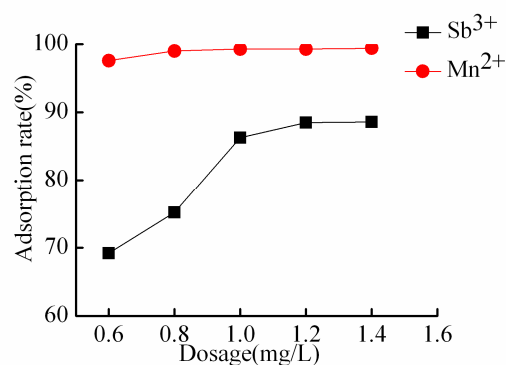


Figure 5. Effect of dosage on adsorption

3.2.2 Effect of contact time on adsorption capacity and adsorption rate

The effect of contact time on adsorption was investigated. The contact time ranged from 1 h up to 5 h was selected to investigate the contact time (Figure 6). As shown in the adsorption of Sb^{3+} when the contact time was increased, the adsorption capacity and removal rate was increased, especially within 4 h, adsorption and removal efficiency was increased sharply. When the time was 4 h, the removal rate reached 86.8%, while continuing to extend the contact time to 5 h, the adsorption capacity reaches up to 88.3%. Between 4-5 h, the adsorption capacity increased slowly. That indicated that the adsorption process was the fast adsorption process, when the concentration of Sb^{3+} was higher, the more adsorbent exposure, the more opportunities for the Sb^{3+} contact with the adsorbent, with the extension of adsorption time, the concentration of the Sb^{3+} becoming lower and the chance of molecular adsorbent contact with the Sb^{3+} reduced, resulting in the adsorption rate decreased. So, the 4 h was selected as the optimum time. But the adsorption of Mn^{2+} was a fast progress, the adsorption rate could reach 96.4% when the contact time was 1 h, extend the contact time to 3 h, it could reach up to 98.9%, continuing extend the contact time to 5 h, the adsorption rate increased smoothly, so 3 h was selected as the optimum time when using NaOH modified zeolite to adsorb the Mn^{2+} .

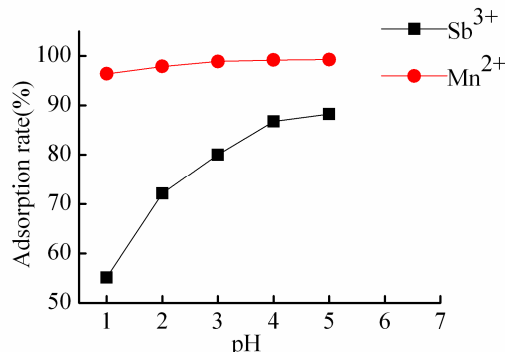


Figure 6. Effect of contact time on adsorption

3.2.3 Effect of Sb^{3+} and Mn^{2+} initial concentration on adsorption capacity and adsorption rate

The effect of the Sb^{3+} concentration on modified zeolite adsorption was investigated. Sb^{3+} concentration ranged from 6.0 mg/L up to 14.0 mg/L was selected to investigate the adsorption effect (Figure 7). As shown in as the initial concentration of Sb^{3+} in solution increased, the removal rate of the modified zeolite of the Sb^{3+} from the initial concentration of 6.0 mg/L of 49.3% rising to 10.0 mg/L of 85.4%, and then gradually decreased to 14.0 mg/L of 80.1%. So, the 10.0 mg/L was selected as the optimum initial concentration. The same principle the 100.0 mg/L Mn^{2+} was selected as the optimum initial concentration.

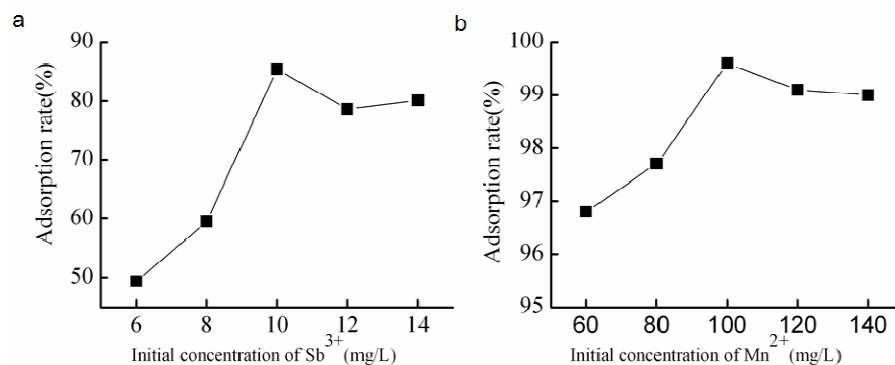


Figure 7. Effect of initial concentration on adsorption

3.2.4 Effect of temperature on adsorption capacity and adsorption rate

The effect of the temperature on modified zeolite adsorption was investigated. The temperature was ranged from 25 up to 45 °C was selected to investigate the adsorption effect (Figure 8). As shown in with the increased of the temperature, the adsorption efficiency of Sb³⁺ and adsorption capacity increased firstly and then decreased when the temperature exceed 35 °C. When the temperature increased from 25 °C to 35 °C, the adsorption efficiency increased from 85.6% to 89.1%. That because as the initial temperature of the solution increased, the movement of Sb³⁺ increased, which lead to accelerate the opportunity of their mutual collisions, and resulting in increased adsorption capacity and adsorption efficiency, but when the temperature exceed 35 °C, the desorption rate on the surface accelerate, which lead to the removal rate decreased to 79.3%. But the temperature from 25 °C to 35 °C on the adsorption of Mn²⁺ was not obvious, the adsorption rate were all exceeded to 99.0%.

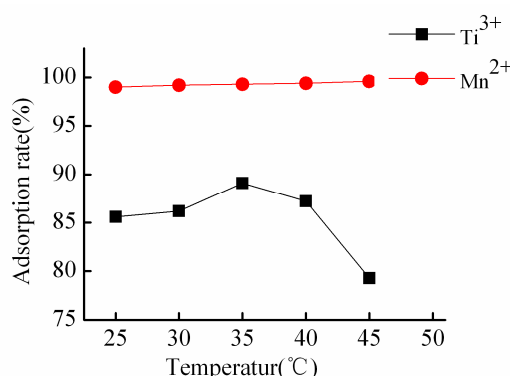


Figure 8. Effect of temperature on adsorption

3.3 Adsorption isotherm

At room temperature, the initial concentration of Sb³⁺ were 5.0 mg/L, 10.0 mg/L, 15.0 mg/L, 20.0 mg/L, 25.0 mg/L, respectively, at the condition of contact time was 4 h, measured the residual concentration of Sb³⁺ in solution when reached the experimental adsorption equilibrium point, then calculated adsorption capacity, Langmuir and Freundlich equations were used to fit the data, the results shown in Figure 6(a) and Figure 6(b) which was the modified zeolite adsorption isotherm of Sb³⁺, the adsorption capacity of Sb³⁺ by HNO₃ modified zeolite increased with the increased of initial concentration, and gradually become saturated. Which due to the larger initial concentration, the greater the driving force for absorption, resulting in increased adsorption capacity. The Langmuir equation correlation coefficient was 0.8198, the Freundlich equation correlation coefficient was 0.9120, which indicated the isotherm process was well predicted by Freundlich model. And the initial concentration of Mn²⁺ were 50.0 mg/L, 100.0 mg/L, 150.0 mg/L, 200.0 mg/L, 250.0 mg/L, respectively, the results shown in Figure 6(c) and Figure 6(d) which was the modified zeolite adsorption isotherm of Mn²⁺, the adsorption capacity of Mn²⁺ by NaOH modified zeolite increased with the increased of initial concentration, and gradually become saturated. Which due to the larger initial concentration, the greater the driving force for absorption, resulting in increased adsorption capacity. The Langmuir equation correlation coefficient was 0.9805, the Freundlich equation correlation coefficient was 0.9386, which indicated the Langmuir equation fitting resulted better.

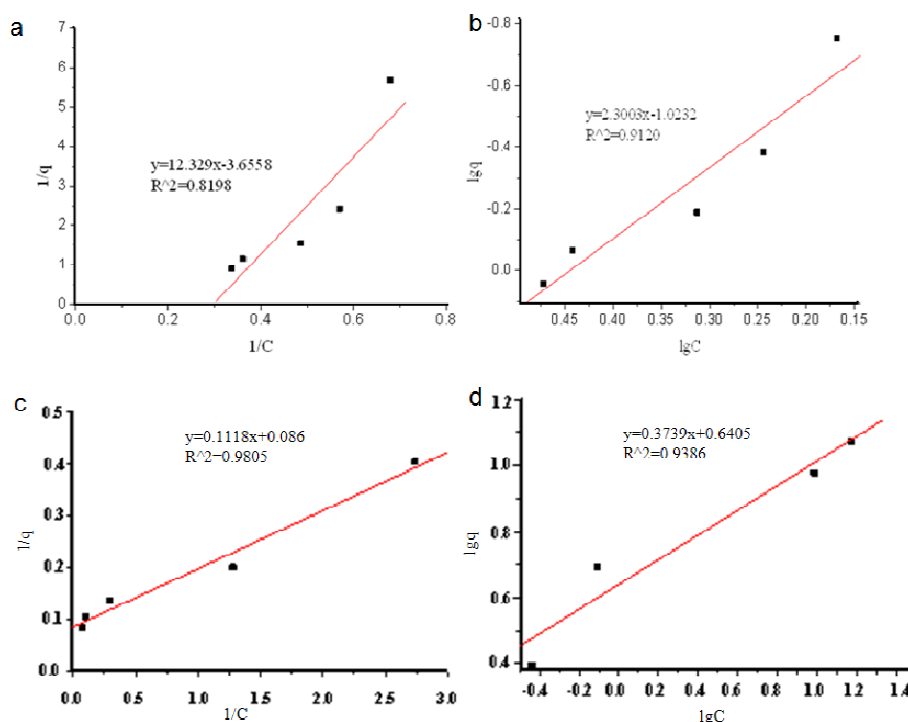


Figure 9. Adsorption isotherm of Sb^{3+} , Mn^{2+} fitted to Langmuir model and Freundlich model

3.4 Adsorption kinetics

The Sb^{3+} adsorption kinetics on modified zeolite was studied by using Sb^{3+} initial concentration of 10.0 mg/L. As shown in Figure 10 that the Sb^{3+} and Mn^{2+} adsorption capacity also increased with increasing contact time. Two commonly used kinetic models, pseudo first-order and second-order kinetic models have been applied to describe the adsorption of Sb^{3+} and Mn^{2+} onto modified zeolite as a function of contact time. The pseudo first-order kinetic model is expressed as below:

$$\ln(q_e - q_t) = \ln q_e - k_1 t$$

where, q_t (mg/g) and q_e (mg/g) represent the amount of adsorbate adsorbed at time t and at equilibrium time, respectively, and k_1 represents the adsorption rate constant. The adsorption rate constant (k_1) was calculated from the plot of $\ln(q_e - q_t)$ against t . The pseudo second-order kinetic model can be expressed as below.

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e}$$

where, k_2 (g/(mg·min)) is the pseudo second-order rate constant of sorption, q_e (mg/g) is the amount of adsorbate sorbed at equilibrium and q_t (mg/g) is the amount of adsorbate on the surface of the adsorbent at any time t . The q_e and k_2 can be obtained by linear plot of t/q_t versus t . According to the calculated kinetic parameters shown in Figure 10, it can be concluded that the pseudo second-order kinetic model can produce better fitting to the experimental data of Sb^{3+} and Mn^{2+} adsorption. The result revealed that the chemisorption is significant in the rate controlling step.

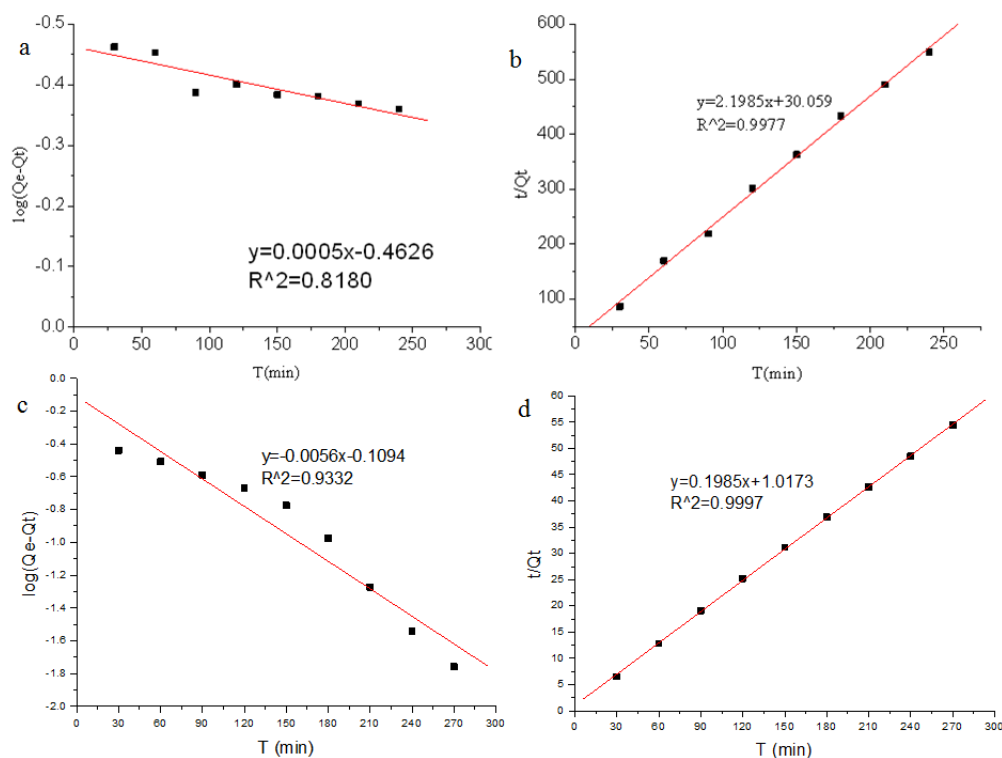


Figure 7. (a)Pseudo first-order kinetic model of Sb^{3+} , (b)Pseudo second-order kinetic model of Sb^{3+} , (c)Pseudo first-order kinetic model of Mn^{2+} , (d)Pseudo second-order kinetic model of Mn^{2+}

CONCLUSION

In this study, the adsorption test for removing Sb^{3+} and Mn^{2+} by modified zeolite have been researched. The single influential factor was investigated, including the dosage of the modified zeolite, the contact time, the initial concentration and the temperature of the solution. The highest adsorption rate of Sb^{3+} and Mn^{2+} were 89.1% and 99.6%, respectively. The isotherm process was well predicted by Freundlich model when adsorption Sb^{3+} by using HNO_3 modified zeolite and the Langmuir model fitting the equation well when adsorption Mn^{2+} by using NaOH . The kinetic process was well predicted by pseudo second-order model when both adsorption Sb^{3+} by using HNO_3 modified zeolite and adsorption Mn^{2+} by using NaOH modified zeolite.

Acknowledgments

This research is supported by Scientific Research Fund of Hunan Provincial Education Department(14B059), Hunan University of Science and Technology Research Project(E51395).

REFERENCES

- [1] X Liang; YH Chen; XF Wu; FY Zhang. *J. Centr. Sou. Univer. Fores. & Techno.*, **2012**, 32(12), 104-109.
- [2] CX Gao; C Li; J Peng; D Su. *J. Chongqing Techno. & Busine. Univer.(Natu. Sci. Edit.)*, **2013**, 30(8):78-83
- [3] W Shoty; B Chen; M Krachler. *J. Environ. Monitor.*, **2005**, 7(12): 1148-1154.
- [4] P Smichowski. *Talanta*, **2008**, 75(1), 2-14.
- [5] M Karniba; A Kabbani; H Holail; Z Olama. *Energy Procedia*, **2014**, 50, 113-120.
- [6] RS Juang; SH Lin; TY Wang. *Chemosphere*, **2003**, 53(10), 1221-1228.
- [7] AS Mohammad; AQ Zakaria; BM Khalid; RQ Mohammed; A Malek. *Chem. Eng. J.*, **2015**, 260, 749-756.
- [8] Q Weng; ZG Zhang. *Chinese Journal of Environmental Engineering*, **2014**, 8(2), 548-552.
- [9] YS Ok; JE Yang; YS Zhang; SJ Kim; DY Chung. *J. Hazard. Mater.*, **2007**, 147(1), 91-96.
- [10] P Zhang; GC Zhu; BZ Ren; YJ Zhang; W Luo. *Journal of Chemical and Pharmaceutical Research*, **2014**, 6(9), 37-42.